National-scale down-the-drain environmental risk assessment of oxybenzone in the United States

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Supplementary Information

Figure S1. Flow chart outlining freshwater environmental risk assessment (ERA) approach used in this paper. (A) grams/per capita/day (g/c/d); (B) see Table S3 for iSTREEM® input parameters; (C) see Table 2 (main text) for 90th percentile PEC; (D) see Table S4 for identified freshwater toxicity data; (E) see Table 1 (main text) for most sensitive data points and PNEC; (G) See Table S5 for determining the appropriate assessment factor. (F) Refinement of the assessment is required (e.g., collection of more toxicity data or refinement of exposure characterization assumptions) when the PNEC exceeds the PEC. Predicted environmental concentration (PEC); predicted noeffect concentration (PNEC).

Figure S2. The discharge locations of the 13 245 wastewater treatment plants accounted for in the iSTREEM[®] model.

Table S1. Summary of BP-3 physico-chemical properties.

Table S1. Summary of wastewater treatment removal data from the literature. The mean value across the 22 values in this table was used for the iSTREEM® modeling. The search was limited to activated sludge treatment as this forms the majority of wastewater treatment plants (WWTPs) in the United States.

Table S3: Summary of inputs to iSTREEM® V2.2 to estimate U.S. national mean-flow PECs.

Table S4. Summary of BP-3 aquatic toxicity to standard freshwater test organisms. Hazard characterization completed by application of an appropriate assessment factor to the most sensitive endpoint identified. Only data related to ecologically relevant endpoints (e.g., mortality, growth, reproduction) from the studies conducted according to or similar to standardized test guidelines were selected for use. Greater than values were not considered to be most sensitive when other data was available. Measured values were always selected in preference to nominal values when

available. *Italicized* results were not considered in the assessment. **Bold** results were selected as most sensitive within the trophic level.

Table S5. Summary of assessment factors to derive predicted no-effect concentrations (PNECs) in this study. These assessment factors are based on U.S. Environmental Protection Agency guidance (Nabholz 1991; Zeeman and Gilford, 1993; Beasley et al. 2018).

Table S6: Global monitoring data collected from literature for BP-3 in WWTP effluent and rivers. Data presentation varied by study. In Figure 2 (in the main text) the mean/median was used for distribution calculations. When data was provided for different rivers or WWTPs or at different timepoints, it was not combined into a single value. In certain cases, data needed to be estimated from a graph. This is reported in the 'Notes' column.

Table S7. Global BP-3 monitoring data in lakes or reservoirs collected from the literature. Data presentation varied by study; the mean/median was used in distribution calculations. When data was provided for different locations or at different timepoints, it was not combined into a single value. In certain cases, data needed to be estimated from a graph.¹.

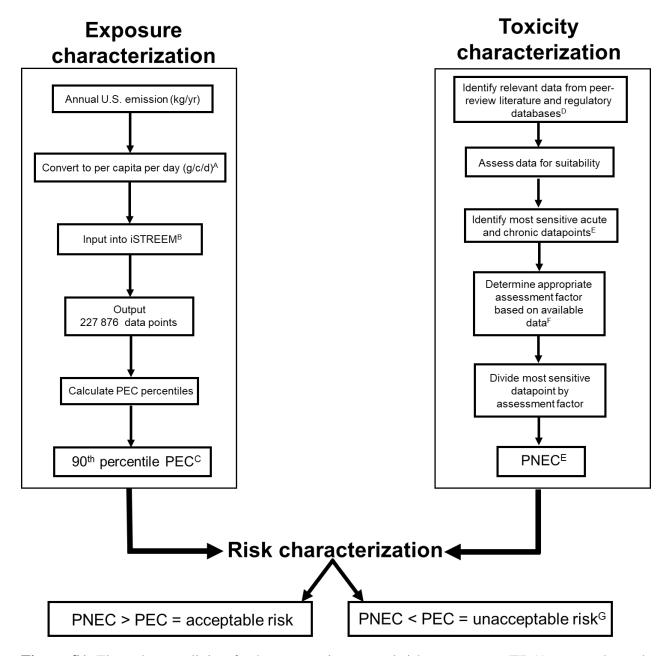


Figure S1. Flow chart outlining freshwater environmental risk assessment (ERA) approach used in this paper. (A) grams/per capita/day (g/c/d); (B) see Table S3 for iSTREEM® input parameters; (C) see Table 2 (main text) for 90th percentile PEC; (D) see Table S4 for identified freshwater toxicity data; (E) see Table 1 (main text) for most sensitive data points and PNEC; (F) See Table S5 for determining the appropriate assessment factor. (G) Refinement of the assessment is required (e.g., collection of more toxicity data or refinement of exposure characterization assumptions) when the PNEC exceeds the PEC. Predicted environmental concentration (PEC); predicted noeffect concentration (PNEC).

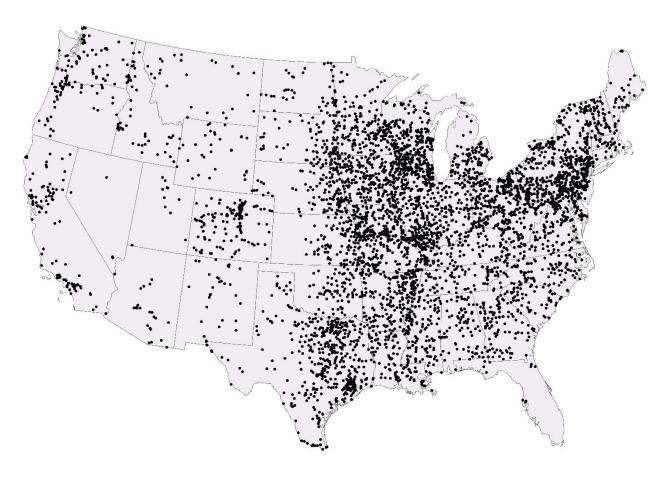


Figure S2. The discharge locations of the 13 245 wastewater treatment plants accounted for in the iSTREEM® model.

Table S1. Summary of oxybenzone (benzophenone-3) physico-chemical properties.

INCI Name (CAS)	Abbreviation	Structure	Solubility ¹	pKa	Log Kow ¹
Benzophenone-3 (131-57-7)	BP-3	O CH ₃	6 mg/L	7.12	3.45

¹ECHA (2020)

Table S2. Summary of data for the removal of oxybenzone (BP-3) by wastewater treatment studies published in the literature. The mean value across the 22 values in this table was used for the iSTREEM® modelling. The search was limited to activated sludge treatment as this forms the majority of wastewater treatment plants (WWTPs) in the United States.

Mean removal (standard deviation)	Mean removal value(s) or range reported	Reference
86% (12)	97%1	Kasprzyk-Hordern et al. (2009)
	82% 1	Bueno et al. (2012)
	94% ¹ (April), 85% ¹ (October) ²	Liu et al. (2012)
	60% (winter), 62% (summer)	Golovko et al. (2014)
	83% (dry season) ¹ , 92% (wet season) ¹	Tsui et al. (2014a)
	87% 1	Ekpeghere et al. (2016)
	86.2%, 91.7%	Wang and Kannan (2017)
	68%, 93%, 93%, 96%, 74%, 99%, 92%, 98%, 99%	Balmer et al. (2005)
	78.2%	Rosal et al. (2010)

¹Values estimated from a graph using WebPlot Digitizer (Rohatgi 2020).

²Predicted, National Center for Biotechnology Information (2020)

²Removal values reported separately for primary treatment followed by activated sludge. These removals were added together for consistency with other reported values.

Table S3: Summary of inputs to iSTREEM® V2.2 to estimate U.S. national mean-flow PECs.

Parameter	Input
WWTP removal for: Activated sludge, oxidation ditch, lagoon, trickling filter, and rotating biological contractor	86%
Primary removal	0%
In-stream decay (d ⁻¹)	$0.014~{\rm d}^{-1}$
Emission	0.011 g/c/d

Table S4. Summary of BP-3 aquatic toxicity to standard freshwater test organisms. Hazard characterization completed by application of an appropriate assessment factor to the most sensitive endpoint identified. Only data related to ecologically relevant endpoints (e.g., mortality, growth, reproduction) from the studies conducted according to or similar to standardized test guidelines were selected for use. Greater than values were not considered to be most sensitive when other data was available. Measured values were always selected in preference to nominal values when available. Italicized results were not considered in the assessment. **Bold** results were selected as most sensitive within the trophic level.

Species	Life stage	Duration	Endpoint	Result (µg/L)	Measured/ Nominal	Guideline	Source	Year	Conclusion
Oryzias latipes	n/a	96 h	LC50	3800	measured	~OECD 203	ECHA	2005	Valid
Poecilia	Adult	96 h	LC50	> 1	nominal	n/a	Almeida et	2019	Invalid results are
reticulata			NOEC (mortality, condition, ertythrotoxicity)	≥1	nominal	-	al.		non-standard endpoints and/or lack known
			NOEC (genotoxicity)	0.01	nominal				ecological relevance.
			NOEC (genotoxicity)	0.1	nominal	_			
Danio rerio	n/a	96 h	LC50	3900	n/a	GB/T 27861-2011	Du et al.	2017	Valid
Danio rerio	Larvae	6 d	NOEC (gene expression)	< 1	n/a	n/a	Zhang et al.	2017	Non-standard endpoint. Ecological relevance unknown.
Danio rerio	Eleutheroembryo	48 h	NOEC (gene expression)	100	measured	n/a	Rodrigues- Fuentes et al.	2015	Non-standard endpoint. Ecological relevance unknown.
Danio rerio	Embryo	96 hpf	LC50	> 100	nominal	n/a	Tao et al.	2020	Invalid results are
		96 hpf	NOEC (mortality)	≥ 100	nominal	_			non-standard
	Larvae	96 hpf	NOEC (neorotoxicity)	1	nominal	-		lack	endpoints and/or lack known ecological
		11 dpf	NOEC (neurotoxicity)	< 10	nominal	_			relevance.
Danio rerio	Embryo	96 h	LC50	15900	measured	~OECD 236	Balazs et al.	2016	Valid
		12 0h	LC50	13100	measured	_			

Table S4. Continued.

					Result	Measured/				
	Species	Life stage	Duration	Endpoint	(µg/L)	Nominal	Guideline	Source	Year	Conclusion
	Danio rerio	Adult	14 d	LC50	> 312	measured	OECD 204	Bluthgen et	2012	Invalid results are
			14 d	NOEC (mortality, growth, histology)	≥ 312	measured	_	al.		non-standard endpoints or
			14 d	NOEC (gene expression)	< 2.4	measured	_			exposure and/or lack known ecological
		Embryo	120 hpf	LC50	> 438	measured	n/a	_		relevance.
			120 hpf	NOEC (mortality, behavior)	≥ 438	measured	-			
			120 hpf	NOEC (gene expression)	8.2	measured	-			
	Danio rerio	Embryo (F1)	24 h	LC50	> 550	nominal	n/a	Li et al.	2018	Invalid results are
			24 h	NOEC (mortality)	5.5	nominal	_			non-standard endpoints or
			48 h	NOEC (heart rate)	< 5.5	nominal	_			methods and/or
FISH			72 h	NOEC (hatch)	≥ 550	nominal	_			lack known
至		Embryo-larvae	48 h	NOEC (heart rate)	< 0.055	nominal	n/a	_		ecological relevance.
			72 h	NOEC (hatch)	0.055	nominal	_			
			7 d	NOEC (mortality)	≥ 550	nominal	_			
			7 d	LC50	≥ 550	nominal	_			
	Danio rerio	Embryo	6 d	LC50	2000 - 4000	nominal	OECD 212	Jang et al.	2016	Valid
				NOEC (mortality, hatch, defect)	2000	nominal	_			
	Oncorhynchus mykiss	Juvenile	14 d	NOEC (VTG)	80	measured	n/a	Coronado et al.	2008	Invalid results are non-standard
	Oryzias latipes	Adult	21 d	NOEC (VTG)	132	measured	n/a			endpoints and/or
			21 d	NOEC (fecundity)	≥ 620	measured	(~OECD 229)		lack known ecological	
			21 d	NOEC (hatch)	132	measured	/			relevance.

Table S4. Continued.

					Result	Measured/				
	Species	Life stage	Duration	Endpoint	$(\mu g/L)$	Nominal	Guideline	Source	Year	Conclusion
	Oryzias latipes	Adult	14 d	LC50	> 90	measured	n/a	Kim et al.	2014	Invalid results are
			14 d	NOEC (growth,	≥ 90	measured	_			non-standard endpoints or
				mortality, organ						methods and/or
				indices)						lack known
			14 d	NOEC (E2, T,	26	measured	_			ecological
				VTG)			_			relevance.
			14 d	NOEC (gene	8.4	measured				
			-	expression)			_			
			14 d (28 d	NOEC	8.4	measured				
			total)	(reproduction)						
		Embryo-larvae	30 dph	LC50	> 500	nominal	_			
		(F1)	30 dph	NOEC (mortality,	≥ 500	nominal	_			
				hatch, growth)						
H	Pimephales promelas	Juvenile	14 d	LC50	> 3900	measured	n/a	Kunz et al.	2006	Invalid results are
FISH			14 d	NOTC (UTC)	> 2000		_			non-standard
			14 0	NOEC (VTG)	≥ 3900	measured				endpoints and/or lack known
			14 d	NOEC (growth)	766	measured	_			ecological
										relevance. Growth
										endpoint is valid.
	Danio rerio	Embryo-Adult	60 dph	LC50	≥470	measured	OECD 234	Kinnberg et	2015	Invalid results are
			60 dph	NOEC (mortality,	191	measured	_	al.		non-standard
			оо ари	growth)	1,1	111000000100				endpoints, lack appropriate conc
			60 dph	NOEC (sex ratio,	191	measured	_			response and/or
			_	gonad						lack known
				maturation)						ecological
			60 dph	NOEC (VTG)	≥ <i>470</i>	measured				relevance.
		Adult	12 d	LC50	> 437	measured	n/a	_		
			12 d	NOEC (VTG)	63	measured	_			

Table S4. Continued.

					Result	Measured/				
	Species	Life stage	Duration	Endpoint	$(\mu g/L)$	Nominal	Guideline	Source	Year	Conclusion
	Carassius auratus	n/a	28 d	LC50	> 5	nominal	n/a	Liu et al.	2015b	Non-standard test species. Invalid
			28 d	NOEC (mortality)	> 5	nominal				results are non- standard endpoints
H			28 d	NOEC (hepatic oxidative stress)	0.5	nominal	-			and/or lack known ecological relevance.
FISH	Betta splendens	Adult	28 d	LC50	> 553	measured	n/a	Chen et al.	2016	Non-standard test
_			28 d	NOEC (GSI, mortality)	≥ 553	measured	-			species. Invalid results are non-
			28 d	NOEC (behavior)	< 5.5	measured	_			standard endpoints and/or lack known
			28 d	NOEC (histology)	55	measured	-			ecological relevance.
	Chironomus riparius	Fourth instar larvae	24 h	LC50	> 10000	nominal	n/a	Ozaez et al.	2013	Invalid results are non-standard endpoints and/or
			24 h	NOEC (gene expression, mortality)	1000	nominal				lack known ecological relevance.
闰	Chironomus riparius	Embryo	24 h	NOEC (gene expression)	< 910	n/a	n/a -	Ozaez et al.	2014	Invalid results are non-standard
ZAT			72 h	EC50 (hatch)	< 910	n/a				endpoints and/or lack known
EBI			72 h	NOEC (hatch)	< 910	n/a				ecological relevance.
INVERTEBRATE			> 72 h (hatch)	LC50	> 910	n/a				Televanee.
IN			> 72 h (hatch)	NOEC (mortality)	> 910	n/a	-			
	Chironomus riparius	Embryo	24 h	NOEC (gene expression)	< 100	nominal	n/a	Ozaez et al.	2016	Invalid results are non-standard
	•	Fourth instar larvae	24 h	NOEC (gene expression)	100	nominal	-			endpoints and/or lack known ecological relevance.

Table S4. Continued.

S	Species	Life stage	Duration	Endpoint	Result (µg/L)	Measured/ Nominal	Guideline	Source	Year	Conclusion
C	Thironomus iparius	Fourth instar larvae	8 h	NOEC (gene expression)	< 100	nominal	n/a	Martin- Folgar et al.	2018	Invalid results are non-standard endpoints and/or
			24 h	NOEC (gene expression)	< 100	nominal	_			lack known ecological relevance.
	Chironomus riparius	Fourth instar larvae	8 h	NOEC (gene expression)	< 100	nominal	n/a	Martinez- Guitarte	2018	Invalid results are non-standard
			24 h	NOEC (gene expression)	100	nominal	_			endpoints and/or lack known ecological relevance.
	Chironomus iparius	Fourth instar larvae	8 h	NOEC (gene expression)	< 10	nominal	n/a	Muniz- 202 Gonzalez and Martinez- Guitarte	2020	Invalid results are non-standard
			8 h	NOEC (enzymatic activity	≥ 100	nominal	_			endpoints and/or lack known ecological
			24 h	NOEC (gene expression)	< 10	nominal	_	Guitarte		relevance.
			24 h	NOEC (enzymatic activity)	10	nominal	_			
-			96 h	LC50	> 10000	nominal	_			
4			96 h	NOEC (mortality)	≤10	nominal	_			
D	Daphnia magna	<24h neonate	48 h	EC50 (mobility)	1870	measured	~OECD 202	ЕСНА	2005	Valid
\overline{D}	Daphnia magna	<24h neonate	48 h	EC50 (mobility)	1900	n/a	OECD 202	Fent et al.	2010	Valid
D	Daphnia magna	<8h neonate	48 h	EC50 (mobility)	2170	nominal	OECD 202	Jang et al.	2016	Valid
D	Daphnia magna	6-24h neonate	48 h	EC50 (mobility)	2010	n/a	OECD 202	Liu et al.	2015a	Valid
D	Daphnia magna	6-24h neonate	48 h	EC50 (mobility)	1100	nominal	GB/T 16125-2012	Du et al.	2017	Valid
D	Daphnia magna	neonate	48 h	EC50 (mobility)	1900	nominal	ISO 6341	Molins- Delgado et al.	2016	Valid

Table S4. Continued.

					Result	Measured/				
	Species	Life stage	Duration	Endpoint	$(\mu g/L)$	Nominal	Guideline	Source	Year	Conclusion
	Dugesia japonica	n/a	48 h	LC50	900	nominal	n/a	Li et al.	2012	Non-standard test species without
			96 h	LC50	500	nominal	_			guideline. Additionally, species native to SE Asia and thus not relevant for U.S. ERA.
	Daphnia magna	< 24h neonate	48 h	EC50 (mobility)	1200	nominal	~OECD 211	Boyd et al.	2021	Deviations from guideline. Wide
INVERTEBRATE			21 d	NOEC (mortality)	2	nominal	-			and limited exposure concentration selection limits 21d endpoint utility.
Z	Daphnia magna	< 24h neonate	48 h	EC50 (mobility)	1670	nominal	OECD 202	Sieratowicz	2011	Valid
Z			21 d	NOEC (mortality, reproduction, growth)	≥ 342	measured	OECD 211	- et al.		
	Daphnia magna	neonate	24 h	24h NOEC (feeding)	400	nominal	n/a	Pablos et al.	2015	Invalid endpoint is not standard.
		< 24h neonate	21 d	21d NOEC (mortality, reproduction, growth)	≥ 200	nominal	OECD 211	-		
ले	Raphidocelis subcapitata	n/a	72 h	EC50 (growth rate)	670	measured	~OECD 201	ЕСНА	2005	Valid
ALGAE/			72 h	NOEC (growth rate)	180	measured	-			
A	Chlorella	n/a	72 h	EC50 (growth)	22400	nominal	~OECD 201	Pablos et al.	2015	Valid
	vulgaris		72 h	EC10 (growth)	2200	nominal				
	·	-	·	<u> </u>		·	·	·		

Table S4. Continued.

					Result	Measured/				
$S_{]}$	pecies	Life stage	Duration	Endpoint	$(\mu g/L)$	Nominal	Guideline	Source	Year	Conclusion
	Desmodesmus ubspicatus	n/a	72 h	EC50 (growth rate)	960	nominal	OECD 201	Sieratowicz et al.	2011	Valid
	•		72 h	EC10 (growth rate)	610	nominal	_			
	Chlamydomonas einhardtii	n/a	72 h	EC50 (growth rate)	2290	nominal	OECD 201	Mao et al.	2018	Non-standard test species.
-	Chlorella sp	n/a	72 h	EC50 (dry weight)	> 11400	nominal	n/a	Zhong et al.	2019	Non-standard test species and
A	Arthrospira sp.	n/a	72 h	NOEC (dry weight)	228	nominal	_			methods, see Burns and Davies
5AC.			72 h	EC50 (dry weight)	2280 - 11400	nominal	_			(2020) for further details.
			72 h	NOEC (dry weight)	228	nominal	_			
_	cenedesmus bliquus	n/a	96h	EC50 (growth rate)	3640	nominal	n/a	Lee et al.	2020	Non-standard test species without guideline.
	Chlorella ulgaris	n/a	96h	EC50 (growth rate)	2980	nominal	GB/T 21805-2008	Du et al.	2017	Valid
C	Chlamydomonas einhardtii	n/a	96h	EC50 (growth rate)	5000	nominal	n/a	Esperanza et al.	2019	Non-standard test species without
			96h	NOEC (growth rate)	1250	nominal	_			guideline.
	Aicrocystis eruginosa	n/a	10d	EC50 (growth rate)	>5000	nominal	n/a	Mao et al.	2017	Non-standard test species and
	teruginosa Chlamydomonas ceinhardtii	n/a	10d	EC50 (growth rate)	1850	nominal	_			methods.

Table S5. Summary of assessment factors to derive predicted no-effect concentrations (PNECs) in this study. These assessment factors are based on U.S. Environmental Protection Agency guidance (Nabholz 1991; Zeeman and Gilford, 1993; Beasley et al. 2018).

Assessment factor
1000
1000
100
10
1

¹To apply the chronic data assessment factor (i.e. 10), it must be applied to the most sensitive taxa. The acute toxicity profile can be useful for this purpose.

Table S6. Global monitoring data collected from literature for BP-3 in WWTP effluent and rivers. Data presentation varied by study. In Figure 2 (in the main text) the mean/median was used for distribution calculations. When data was provided for different rivers or WWTPs or at different timepoints, it was not combined into a single value. In certain cases, data needed to be estimated from a graph. This is reported in the 'Notes' column.

	M	latrix (μg/L)			
Country	Effluent	River	Result type	Notes	Reference
Switzerland	0.3		Mean	Collected at Horgen WWTP in April, 2002.	Balmer et al.
	0.05		Mean	Collected at Meilen WWTP in April, 2002.	(2005)
	0.065		Mean	Collected at Kusnacht WWTP in April, 2002.	
	0.15		Mean	Collected at Männedorf WWTP in April, 2002.	
	0.6		Mean	Collected at Meilen WWTP in June, 2002.	
	0.3		Mean	Collected at Thalwil WWTP in June, 2002.	
	0.7		Mean	Collected at Wädenswil WWTP in June, 2002.	
	0.01		Mean	Collected at Wetzikon WWTP in June, 2003.	
	0.2		Mean	Collected at Männedorf WWTP in Sept., 2003.	
	0.02		Mean	Collected at Thalwil WWTP in Sept., 2003.	
	< 0.01		Mean	Collected at Kloten WWTP in Sept., 2003.	
		0.062	Mean, n=2	Samples collected from River Glatt. POCIS were	Fent et al.
				deployed but only results from grab samples were reported. Grab samples collected twice in 2007.	(2010b)
Germany		0.008 (0.007 – 0.011)	Median (range)	Rivers discharging into Baltic Sea. Sampled twice in 2015 at 5 different stations.	Fisch et al. (2017)
	0.431 (0.043)		Mean (SD)	Samples collected for the purpose of method development. Location not provided.	Moeder et al. (2010)
	0.045 (0.005)		Mean (SD)	Collected from WWTP in Leipzig.	Rodil et al. (2009)
	< LOD		Mean (SD)	River Elsterbecken, May 2007.	Rodil and
	0.03 (0.003)		Mean (SD)	River Parth, May 2007.	Moeder (2008)

Table S6. Continued.

	Matrix (μg/L)				
Country	Effluent	River	Result type	Notes	Reference
Germany	0.096 (0.012)		Mean (SD)	WWTP 1	Wick et al.
•	< LOQ		Mean (SD)	WWTP 2; LOQ =25 ng/L.	(2010)
		< LOQ	Mean (SD)	River Rhine; $LOQ = 5 \text{ ng/L}$.	
		< LOQ	Mean (SD)	Stream 1; $LOQ = 5 \text{ ng/L}$.	
		0.047 (0.029)	Mean (SD)	Stream 2.	
Slovenia	0.015 (0.002 – 0.049)	0.007 (0.005 -0.044)	Median (range)	Wastewater treatment plants sampled in Slovenia and Croatia (n=12 samples) The Sava River was sampled at seven locations twice (n=14), including upstream of wastewater discharges.	Česen et al. (2019)
Spain	0.016 (0.008 – 0.034)	n.d. (n.d. – 0.038)	Median (range)	Five river water samples were collected at various points long upper and lower point of Llobregat River in October 2011 Five WWTPs in the Catalonia region sampled October 2011.	Gago-Ferrero et al. (2013)
				Samples were collected from WWTP influent, effluent and surface water (river) from June – September 2008. Samples collected during the highest consumption season (UV filters) with lowest flow rate.	Negreira et al. (2009)
	< LOD	0.054 (0.003)	Mean (SD)	June 2008 sampling. $LOD = 13 \text{ ng/L}$.	
	0.083 (0.012)	0.087 (0.008)	Mean (SD)	July 2008 sampling.	
	0.077 (0.004)	< LOD	Mean (SD)	August 2008 sampling. BP-3 LOD = 8 ng/L .	
	0.084 (0.003)	< LOD	Mean (SD)	September 2008 sampling. BP-3 LOD = 8 ng/L.	
				Samples collected in Catalonia.	Pedrouzo et
	n.d.		(n=3)	WWTP A; $BP-3 LOD = 5 ng/L$.	al. (2010)
	< LOQ		(n=3)	WWTP B; BP-3 LOD = 5 ng/L .	
	< LOQ		(n=3)	WWTP C; BP-3 LOD = 5 ng/L .	
	< LOQ		(n=3)	WWTP D; BP-3 LOD = 5 ng/L .	

Table S6. Continued

Matrix (μg/L)					
Country	Effluent	River	Result type	Notes	Reference
Spain		0.006 0.008 0.028	(n=3) (n=3) (n=3)	Ebro River Ter River Llobregat River	Pedrouzo et al. (2010)
	0.094 (0.048 – 0.14)		Mean (range)	Monitored effluent in 5 WWTPS in Spain over two years. Mean calculated from range.	Bueno et al. (2012)
		$0.034 \; (0.025 - 0.058)$	Median (range)	Three samples collected in the Besòs River during May, December and July.	Serra-Roig et al. (2016)
		0.035 (0.002)	Mean (SD)	Samples collected in December 2010. From Guadelete River.	Corada- Fernández et al. (2017)
	0.086 (< LOQ – 0.121)		Mean (range)	Wastewater samples were taken every month from a STP located in Madrid (input and output of the secondary clarifier).	Rosal et al. (2010)
Portugal	0.035 (n.d. – 0.068)		Mean (range)	Values are reported in the text only. Samples collected from 15 different wastewater treatment plants.	Cunha et al. (2015)
Italy				Samples collected from alpine river, Adige River near a tourist area. Values estimated from graph and median taken.	Mandaric et al. (2017)
		0.053 (1.86)	Median (SD)	Summer.	
		0.012 (0.005)	Median (SD)	Winter.	
				Samples collected from 8 municipal wastewater treatment plants in Genoa.	Magi et al. (2013)
	0.005		Mean (SD)	WWTP influent April. Only mean reported for effluent.	
	$0.015 \; (0.008 - 0.028)$		Mean (SD)	WWTP influent May.	
	$0.008 \; (0.005 - 0.010)$		Mean (SD)	WWTP influent June.	
	0.019 (0.010 - 0.026)		Mean (SD)	WWTP influent July.	
	0.013		Mean (SD)	WWTP influent August. Only mean reported for effluent.	
	0.008 (0.006 – 0.01)		Mean (SD)	WWTP influent September.	

Table S6. Continued.

	Matrix (μg/L)				
Country	Effluent	River	Result type	Notes	Reference
Wales	0.22 (< 0.08 – 0.223)	< 0.015	Mean (range)	Results from River Ely at Talbot Green and WWTP Coslech. LOD reported as 15 ng/L. Note there is a typo in the Table it says μ g/L, but refers to ng/L.	Kasprzyk- Hordern et al. (2009)
		< 0.015	Mean	River Ely at Perterson-Super-Ely.	
	$0.231 \ (< 0.08 - 2.2)$	0.009 (< 0.015 -0.043)	Mean (range)	Results from River Taff at Abercynon and WWTP Cilfynydd.	
		0.01 (< 0.015 - 0.044)	Mean (range)	River Taff at Pontypridd.	
Norway				Samples collected from two WWTPs in Norway. Daily concentrations were estimated from the graph.	Krzeminski et al. (2017)
	1.10		(n=1)	VEAS Day 1.	
	0.603		(n=1)	VEAS Day 2.	
	0.183		(n=1)	Bekkelaget 1.	
	0.172		(n=1)	Bekkelaget 2.	
	0.143		(n=1)	Bekkelaget 3.	
	0.853 (0.607 -1.10)		Mean (range)	Samples collected from VEAS in WWTP in 2014.	Thomas et al. (2014)
	0.293 (0.081 – 0.598)		Mean (range)	Samples collected from Norway's largest WWTP, VEAS in 2013.	Langford et al. (2015)
	0.233(0.01-0.438)		Mean (range)	Smaller WWTP, HIAS.	
	0.721 (0.374 – 1.92)		Mean (range)	Smaller WWTP Thomasjord.	
USA				Samples collected from two WWTPs in New York state. 24-h composites collected consecutively over a 7-day period.	Wang and Kannan (2017)
	<loq (loq="" –<br="">0.012)</loq>		Median (SD)	WWTP A, 15 000 population served. LOQ= 0.5 ng/L.	
	<loq (loq="" –<br="">0.033)</loq>		Median (SD)	WWTP B, 100 000 population served.	

Table S6. Continued.

N		atrix (µg/L)			
Country	Effluent	River	Result type	Notes	References
USA				Samples collected upstream and downstream of recreational activity in Clear Creek at Golden, Colorado. Samples were collected hourly over a 72-h holiday weekend (Labor Day). Based on their data, they predicted a recreational release of 122 ng/L. All values estimated from graphs.	Reed et al. (2017)
		$< 0.05 \ (< 0.05 - 0.09)$	Mean (range)	Upstream ($n = 72$). There was a 1% detection rate.	
		$0.07 \ (< 0.05 - 0.72)$	Mean (range)	Downstream ($n = 72$). There was a 30% detection rate.	
		0.026 - 0.325	Range	Samples collected at three locations in Clear Creek at Golden in Colorado. Mean cannot be calculated as it is unclear how many samples were <lod. distribution<="" from="" mec="" omitted="" td=""><td>Rand et al. (2020)</td></lod.>	Rand et al. (2020)
				Samples collected from three southwestern rivers over Labor Day weekend, 2017. Samples were collected from an upstream and downstream site at 8 am and 4 pm Sunday – Wednesday.	Rand et al. (2020)
		$0.22 \ (< 0.03 - 0.86)$	Mean (range)	Salt River, Arizona.	
		$0.06 \ (< 0.03 - 0.13)$	Mean (range)	Truckee River, Nevada.	
		0.25(< 0.03- 0.96)	Mean (range)	Clear Creek, Colorado.	
Australia	0.153 (0.121)		Mean (SD)	Bolivar WWTP in Adelaide, South Australia.	Liu et al. (2012)
	0.033 (0.002)		Mean (SD)	Effluent samples collected from the Bolivar sewage treatment plant in South Australia.	Liu et al. (2011)

Table S6. Continued.

	Matrix (μg/L)		_		
Country	Effluent	River	Result type	Notes	References
Australia				Samples collected in August 2011 from four rivers that drain into Port Philip Bay near Melbourne. Only the mean concentration is provided.	Allinson et al. (2018)
		0.006	Mean	Mouth of Werribee River.	
		0.004	Mean	Mouth of Kororit Creek.	
		0.007	Mean	Yarra River estuary.	
		0.004	Mean	Maribyrnong River prior to confluence with Yarra River.	
China				Samples collected at the influent and effluent of the Tianjin wastewater treatment plant. Includes ozonation as the secondary treatment.	Li et al. (2007)
	0.086 (0.068 -0.103)		Mean (range)	Collected February 2005.	
	0.438 (0.377 – 0.506)		Mean (range)	Collected July 2005.	
	0.226 (0.210 -0.246)		Mean (range)	Collected September 2005.	
		0.010 (0.03)	Mean (max.)	Samples (n=27) were collected along the Huangpu River and the surface water in and near Water Park within Shanghai.	Wu et al. (2017)
Hong Kong	0.111 (0.541)		Mean (SD)	Samples were collected from 5 WWTP equipped with different treatment levels.	Tsui et al. (2014a)
Taiwan	0.021 0.013	0.012	(n=1) (n=1)	Two municipal wastewater treatment plants, effluents were collected in different months in the city of Tainan. Surface water samples came from two major rivers in northern Taiwan (where the effluent would be discharged). WWTP Effluent 1. WWTP Effluent 2.	Wu et al. (2013)
		0.012 0.015	(n=1) (n=1)	River Water 1. River Water 2.	

Table S6. Continued.

	Matrix (μg/L)		_		
Country	Effluent	River	Result type	Notes	References
Japan				Twenty-nine sampling sites in rivers of various anthropogenic influence.	Kameda et al. (2011)
		0.025 (0.016 - 0.041)	Mean (range)	Effluent fed stream, 2 sampling sites.	(- /
		0.004	(n=1)	Heavily polluted rivers, 6 sampling sites.	
		$0.006 \; (0.004 - 0.012)$	Mean (range)	Moderately polluted rivers, 12 sampling sites.	
		$0.004 \ (0.002 - 0.01)$	Mean (range)	Background river, 5 sampling sites.	
	0.054 (0.029 – 0.164)		Mean (range)	WWTP effluents, four sampling points.	
South Korea	0.143 (<lod –<br="">0.502)</lod>	0.009 (<lod 0.156)<="" td="" –=""><td>Median (range)</td><td>Four wastewater treatment plants sampled. Surface water samples were collected from eleven sites along three South Korean rivers. Median concentrations were estimated from graphs. For consistency, sewage treatment data not included.</td><td>Ekpeghere et al. (2016)</td></lod>	Median (range)	Four wastewater treatment plants sampled. Surface water samples were collected from eleven sites along three South Korean rivers. Median concentrations were estimated from graphs. For consistency, sewage treatment data not included.	Ekpeghere et al. (2016)
Thailand		0.101 (0.086 -0.116)	Median (range)	Surface water samples were collected from 8 locations; however, these were all marine surface water with the exception of the Bangkok samples (n=2).	Tsui et al. (2014b)
Singapore				Tropical urban watershed. The main water basin receives discharge from 5 major tributaries.	Mao et al. (2018)
		0.008 (0.002 -1.23)	Median (range)	Tributary.	
				Tropical urban watershed. The main water reservoir receives discharge from 5 major tributaries.	You et al. (2015)
137-1		0.011 (<0.001 – 0.224)	Median (range)	Tributaries.	Carlan LOO

¹Values estimated from a graph using WebPlot Digitizer (Rohatgi 2020). Not detected = n.d.; maximum = max.; limit of detection = LOD; limit of quantification = LOQ.

Table S7. Global BP-3 monitoring data in lakes or reservoirs collected from the literature. Data presentation varied by study; the mean/median was used in distribution calculations. When data was provided for different locations or at different timepoints, it was not combined into a single value. In certain cases, data needed to be estimated from a graph.¹

Country	Lake water (µg/L)	Result type	Notes	Reference
Switzerland	0.016	Mean (n=2)	Collected from Lake Zurichsee, August 2002.	Balmer et al.
	0.011	Mean (n=3)	Collected from Lake Zurichsee, September 2002.	(2005)
	0.020	(n=1)	Collected from Lake Grifensee, August 2002.	
	0.030	(n=1)	Collected from Lake Grifensee, July 2002.	
	0.035	(n=1)	Collected from Lake Huttnersee, July 8, 2002.	
	0.023	(n=1)	Collected from Lake Huttnersee, July 26, 2002.	
Germany	0.083 (0.011)	Mean (SD)	Samples collected from Germany surface water for the purpose of method development. Location not provided.	Moeder et al. (2010)
	0.04 (0.003)	Mean (SD)	Surface water was collected at the Lake Cospuden.	Rodil et al. (2009)
	<lod< td=""><td>Mean (SD)</td><td>Lake Cospudener, May 2007. LOD = 11 ng/L.</td><td>Rodil and Moeder (2008)</td></lod<>	Mean (SD)	Lake Cospudener, May 2007. LOD = 11 ng/L.	Rodil and Moeder (2008)
	0.027 (0.004)	Mean (SD)	Lake Cospudener, June 2007.	
	0.017 (0.002)	Mean (SD)	Lake Bagger, May 2007.	
	0.055 (0.011)	Mean (SD)	Lake Bagger, June 2007.	
Australia	0.646 (0.352 – 0.888)	Median (R)	Samples collected at various time points (12 h) from a freshwater reservoir during swimming. Data estimated from graph.	O'Malley et al. (2021)
France	0.003	Median	Sample collected from Villeneuve-de-la-Raho (artificial lake). This site sees highly increased tourism pressure during the summer months.	Fagervold et al. (2019)
Singapore	0.01 (0.002 – 0.331)	Median (R)	Tropical urban watershed. The main water reservoir receives discharge from 5 major tributaries.	You et al. (2015)
ly,	0.007 (0.005 - 0.056)	Median (R)	Same basin (reservoir) sampled as You et al. (2015).	Mao at al. (2018)

¹Values estimated from a graph using WebPlot Digitizer (Rohatgi 2020).

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